

# Why urban greening requires more than just species biodiversity

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## Abstract

Over the past decades, a noticeable shift has taken place in the relationship between humans and nature. There is a growing interest in re-naturalizing urban environments by increasing vegetation cover as a means to foster more resilient cities and enhance environmental comfort for urban dwellers. Numerous policy recommendations emphasize vague targets such as increasing urban biodiversity or expanding tree canopy cover. While these indicators are indeed relevant and contribute to ecosystem service provision, they remain insufficient as standalone goals, failing to account for the complexity of factors involved in the design and planning of green urban infrastructure. This article places particular emphasis on the role of trees as the core unit of green infrastructure. Other vegetative strata, such as herbaceous layers or vertical greening systems, are acknowledged but regarded as secondary or complementary to tree-based systems. We highlight the conceptual ambiguity surrounding terms like “biodiversity” or “canopy increase” and argue that these metrics, although widely used, must be subordinated to a more integrative and function-driven planning paradigm. Urban forestry strategies should not aim merely to boost biodiversity indices but rather to develop evidence-based planting proposals that consider plant functional traits, optimize ecosystem service delivery, minimize disservices, and ensure species selection aligns with future climate conditions. This also involves maximizing genetic diversity, conducting detailed assessments of below-ground conditions (soils), as well as canopy height and diameter class structures, to enhance resilience against pests and diseases. Furthermore, spatial distribution must be planned to ensure justice and equity across socioeconomic groups. Ultimately, fulfilling all these criteria is likely to result in more biodiverse urban forests, but this diversity will be the consequence of well-defined, objective, and scientifically grounded decisions, rather than an arbitrary increase in a metric whose optimal threshold remains undefined. The goal is not simply more biodiversity, but smarter, functionally informed biodiversity.

**Keywords:** *plant diversity, urban forestry, ecosystem services, climate-resilient cities, green infrastructure*

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## 1. Reintroducing nature into urban spaces: the ecocentric turn

We are witnessing the dawn of a new era, marked by technological advances and shifting paradigms in the human-nature relationship. The relationship between humans and nature has undergone a significant shift in recent decades, moving away from the traditional anthropocentric worldview towards a more ecocentric perspective [1–3]. This shift is reflected in the growing body of research on environmental worldviews and their influence on pro-environmental attitudes and behaviors. Such research is influenced by various factors, including socialization processes, such as education and cultural influences [4, 5], as well as personal values, such as biospheric and altruistic values [6, 7]. The shift towards a more ecocentric worldview has also been observed in various contexts, such as in the field of environmental education [8], sustainable development [9, 10], and environmental policy [11, 12].

Cities, as the most densely populated human settlements, are undergoing a significant paradigm shift toward the reintegration of nature into the urban fabric. This transition is increasingly driven by robust scientific evidence highlighting the multifaceted benefits of urban green spaces, including improvements in physi-

cal and mental health, air quality regulation, and overall urban livability [13–16]. Moreover, the strategic incorporation of nature into urban planning plays a crucial role in climate change adaptation, mitigating the urban heat island effect, and fostering urban biodiversity. These outcomes underscore the pivotal function of nature-based solutions in advancing sustainable, resilient, and inclusive cities [17–19].

Given the proven capacity of urban green spaces and urban green infrastructure more broadly to provide multiple ecosystem services in a measurable and well-documented manner, such as sequestering carbon and producing oxygen through photosynthesis, as well as supporting wildlife via the resulting biomass [18, 20–22]; removing atmospheric pollutants [23–25]; mitigating the urban heat island effect [26–29]; and reducing stormwater runoff [30–34], in addition to delivering diverse social, cultural, and human health benefits [35–38], multiple cities worldwide are increasingly adopting urban greening initiatives (**Table 1**). For a more comprehensive overview of cities globally implementing urban greening strategies, see Kumar et al. [39] and references therein. These strategies are aligned with Sustainable

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Development Goal 11 (SDG 11), which emphasizes making human settlements safe, resilient, and sustainable [40], as well as with complementary strategies such as the 3-30-300 rule. However, global assessments continue to reveal insufficient tree canopy cover in most of the cities analyzed [41–44].

This opinion article examines whether cities should aspire to become both refuges and engines of biodiversity in the Anthropocene and whether such a goal is justified. Drawing on ecological theory, urban resilience, and landscape planning, the article challenges the conventional view of cities as entropy-driven systems and instead highlights their latent potential for ecological regeneration. It calls for a paradigm shift in urban planning, advocating for a multidimensional and integrative framework grounded in functional diversity, long-term resilience, and socio-environmental equity. Rather than pursuing biodiversity as an ab-

stract metric, it proposes a scientifically guided design of urban ecosystems that are adaptive, just, and capable of sustaining ecosystem services in the face of accelerating global change. Although the proposed integrated assessment framework, based on scientific literature review, supports conducting urban forestry interventions in a context-specific manner to enhance tree viability, survival, and ecosystem service provision and would be applicable to any city under any climate given its function of constraining the pool of usable species for each specific location within the urban matrix, it may have greater influence or applicability in Mediterranean-temperate climates, where the authors of this opinion article primarily conduct their research. Nevertheless, we consider that the proposed framework, similar to a multi-criteria analysis, is applicable to any city, though detailed studies are needed to evaluate its applicability in cities with different climates and characteristics.

**Table 1 •** Examples of cities implementing urban greening strategies worldwide. This table presents a non-exhaustive selection of cities demonstrating diverse approaches to urban greening implementation across different geographical contexts, scales, and timeframes. The examples illustrate how multiple cities worldwide are actively considering and implementing urban greening measures as integral components of their sustainable urban development strategies.

City	Country	Timeline	Key objectives	References
Paris	France	2020–2030	Transform 50% of the city into an urban forest; plant 170,000 new trees by 2026	[45, 46]
Barcelona	Spain	2021–2030	Nature-based solutions, ecological connectivity, Tree Master Plan 2017–37	[47, 48]
Milan	Italy	2018–2030	Plant 3 million trees in metropolitan areas for climate neutrality integration	[49, 50]
Singapore	Singapore	Ongoing	Comprehensive urban greening with green corridors and vertical vegetation	[51–53]
Melbourne	Australia	2021–2040	Increase tree canopy cover from 22% to 40%; resilient and diverse forest	[39, 54, 55]
New York	USA	2007–2017	Plant one million trees, focusing on underserved neighborhoods (MillionTreesNYC)	[39, 56]
Copenhagen	Denmark	Ongoing	Green roof policies and street-level greening for flood risk mitigation	[39, 57]
Adelaide	Australia	2018–2035	Increase tree canopy cover from 33% to 40% with water-sensitive design	[39, 58]
Cardiff	UK	2020–2030	10-year afforestation program for carbon neutrality; 82,000+ trees planted	[39, 59]
São Paulo	Brazil	Ongoing	City-wide rain gardens and linear parks for urban drainage management	[39]
Nanjing	China	2018–ongoing	Architectural vertical forests with 800+ trees and 2500 shrubs on towers	[39, 60]

As with the transition from the Gothic to the Renaissance, fueled by the printing press, maritime advances, and global connectivity, our current transition is driven by artificial intelligence, aeronautics, and exploration beyond Earth. Crutzen and Stoermer [61] named this epoch the Anthropocene, characterized by technological acceleration, explosive population growth, and massive resource consumption. Lovelock [62] pushes this further in his concept of the Novacene, where intelligence no longer relies solely on DNA- or RNA-based organisms but on digital systems and “cyborgs.” In both scenarios, the urban habitat becomes central to the human condition, resembling the evolution of eusocial species (Harrison et al., 2018) [63].

*Homo sapiens* has transitioned from a rural to an urban species. The urban ecosystem, despite its tensions and linear metabolism, remains a magnet for human settlement. Urban sprawl, often dictated by market logic and disconnected planning, disrupts ecological equilibrium [64]. Urban areas now cover around 3% of continental surfaces [65], while farmland covers five times more [66]. Forests, home to 80% of terrestrial biodiversity, occupy a shrinking proportion. These disparities underscore humanity’s growing ecological footprint.

Conceptual provocations such as Liam Young’s “Planet City” or the space colonization fantasies of tech billionaires illustrate our ongoing struggle to reconcile rapid urban growth with the imperative of environmental preservation. In this context, the two dominant paradigms of urban development, urban sprawl and compact city models, embody contrasting yet equally imperfect solutions. Each offers distinct advantages and limitations, underscoring the persistent tension between expansion and sustainability, and ultimately revealing their status as compromise strategies rather than definitive answers. (1) Urban sprawl, defined by low-density and decentralized growth, offers benefits such as larger living spaces, access to nature, and reduced local congestion [67, 68]. However, it also leads to greater car dependency, higher emissions, habitat fragmentation, and social segregation [69, 70]. (2) Conversely, compact cities encourage high-density, mixed-use development that enhances energy efficiency, supports sustainable mobility, and helps conserve natural and agricultural land [70–72]. Yet, this model can intensify the urban heat island effect, degrade air quality, and reduce green space availability, impacting public health and biodiversity [17, 26, 73–75].

Despite the paradigm differences between urban development models, the concept of urban resilience is gaining increasing prominence in both academic research and practical implementation, regardless of the specific system or urban development typology. Urban resilience refers to the capacity of cities to absorb, adapt to, and recover from a range of shocks and stresses, including those linked to climate change, natural disasters, and socioeconomic disruptions, while maintaining the continuity of their essential structures and functions [76]. Originally conceptualized within the framework of ecological systems theory [77], resilience has since evolved into a foundational principle of urban design and planning. Recent studies highlight that fostering resilience in urban contexts requires integrated and multidimensional approaches that encompass environmental, social, infrastructural, and institutional components. Such approaches enhance a city’s adaptive capacity and contribute to long-term sustainability. As urban areas are increasingly exposed to more frequent and severe challenges, the development of resilient urban systems has

become a central objective in planning and policy agendas, ensuring their ability to withstand and swiftly recover from adverse events [78].

## 2. Green infrastructure in cities: how do we assess it?

Urban green infrastructure plays a crucial role in strengthening the resilience of cities by delivering a wide array of ecosystem services that support adaptation and recovery in the face of environmental stresses and shocks. Beyond its adaptive functions, green infrastructure also presents significant opportunities for ecological naturalization, contributing to the regeneration of degraded urban ecosystems while simultaneously improving environmental quality and human well-being. This dual capacity to bolster resilience and restore ecological integrity underscores the vital importance of embedding green infrastructure at the core of contemporary urban development strategies [79–83].

From the medieval hortus conclusus to contemporary post-industrial urban parks, cities have progressively incorporated increasingly diverse plant assemblages. However, a widespread misconception, particularly in public discourse and non-scientific areas, is that merely increasing plant biodiversity in urban settings is inherently sufficient. For instance, Andres et al. [84] aimed to clarify the notion of “biodiverse reforestation,” highlighting its relevance for both climate change mitigation and biodiversity conservation. They argue that successful reforestation must deliberately embed biodiversity goals, emphasizing species diversity not as an end in itself but as a strategy to enhance carbon sequestration, support ecosystem functionality, and ensure long-term species survival. This intentionality fosters more resilient and sustainable restoration outcomes.

Although biodiversity is frequently equated with species richness, ecological diversity more accurately captures the structural, functional, and relational complexity of ecosystems [85, 86]. As demonstrated across the literature, the sustained delivery of ecosystem services depends less on the absolute number of species and more on the presence of ecologically functional vegetation and the maintenance of interspecific interactions that uphold critical ecological processes [87–89]. A narrow focus on maximizing species count risks overlooking essential traits, such as drought resistance, phenological complementarity, or pollinator attractiveness, that underpin the ecological performance of urban green spaces. Initiatives such as Miyawaki forests or the “Alboran biodiversity islands” exemplify how strategically designed, functionally diverse plant communities can enhance resilience and generate a broad spectrum of ecosystem services even within high-density urban environments [90, 91].

In a similarly flawed manner, indicators such as the NDVI (Normalized Difference Vegetation Index) and tree canopy cover are widely used to assess urban vegetation, yet they fall short in capturing the ecological complexity underlying the provision and accessibility of ecosystem services. While these metrics estimate the surface presence and spatial distribution of vegetation, they are insufficient to reflect key dimensions such as the functional quality of green spaces, vegetation structure or species diversity. Likewise, canopy cover, limited to a two-dimensional representation, overlooks important factors such as biomass density, vertical

stratification, etc. [92]. Relying exclusively on these indicators risks producing biased or incomplete interpretations of urban greenness.

Although both indicators, diversity and canopy cover, are fundamental components associated with the delivery of urban ecosystem services, efforts to enhance green infrastructure should not rely solely on increasing these metrics at the city scale without critically assessing the specific strategies and contextual conditions under which such increases are implemented. A purely quantitative rise, if detached from spatial, ecological, and socio-functional considerations, risks resulting in interventions that are inefficient, inequitable, or ecologically superficial. In this regard, the insight of Edwin Thompson Jaynes, originally formulated within the realm of probability theory, offers a compelling parallel for urban greening: “*It appears to be a quite general principle that, whenever there is a randomized way of doing something, then there is a nonrandomized way that delivers better performance but requires more thought*” [93]. Applied to the design of urban nature, this principle highlights the necessity of intentional, context-sensitive planning over indiscriminate or purely metric-driven approaches. Consequently, any increase in indicators such as diversity or canopy coverage must be underpinned by a comprehensive understanding of the multiple biotic and abiotic factors that influence the long-term viability of urban vegetation and its capacity to sustain ecosystem service delivery. This is particularly crucial given the growing evidence that a significant proportion of trees planted in urban environments today may not reach maturity or achieve their potential functional value, ultimately compromising the resilience and effectiveness of urban green infrastructure [94].

### 3. Are urban trees the central unit of green infrastructure? Mapping areas for planting and conservation

A fundamental starting point for understanding the needs and requirements associated with urban greening and the development of green infrastructure is the segmentation of green spaces according to vegetation strata, particularly the distinction between herbaceous and woody vegetated areas [95]. In this regard, features such as lawns, green roofs, green façades, and similar forms of green infrastructure should not be considered as providers of ecosystem services beyond their esthetic, visual appeal, habitat provision or the capacity of lawns to absorb water in flood-prone areas [96–99]. The literature indicates that green spaces lacking trees or dense woody vegetation often have negligible or even negative impacts on human well-being [100–103]. This underscores the importance of prioritizing tree-based and multilayered vegetation structures in urban planning if the goal is to achieve meaningful ecological and social benefits. This is not to suggest that other vegetation strata do not offer benefits; however, in spatially constrained urban environments, trees represent a particularly practical and efficient solution. Their vertical structure and large biomass enable them to deliver a disproportionately high return on investment compared to lower vegetation forms, maximizing ecosystem service provision within limited ground space. As such, trees are uniquely positioned to address both ecological and social challenges in dense urban settings.

Considering urban trees as the operational unit of work prior to the multifactorial process described below, a thorough assessment of the existing situation of the urban forest must be carried out before performing any type of planting. In this regard, knowledge of the spatial distribution of trees, identification at the species level, and certain structural characteristics such as height, crown dimensions, and DBH (diameter at breast height) constitute necessary baseline information. This information is necessary for calculating ecosystem services that can serve as a basis for delimiting hotspots and cold spots of provision that need intervention or for developing more complex management methods based on ordination methods using additional information that could be consulted in databases such as TRY [104]. Ultimately, after evaluating all relevant information, and necessarily in a spatial manner, this must be integrated through multi-criteria analysis that allows delimiting those areas of cities with higher priority for planting or for allocating existing economic resources for new urban tree planting [105–111]. Spatial prioritization should constitute the first step in any new urban tree planting project, and among the different selected areas, context-specific interventions and actions should be carried out on each one, considering the multiple factors that the authors propose should be worked on in an integrated manner. Similarly, those areas that already have significant tree cover and are not considered a priority for intervention should also take into consideration the different factors so that management is scientifically guided and effective protection of existing resources is achieved.

### 4. A multidimensional space: what tree factors should be assessed for planting and management?

A fundamental consideration in urban forestry is the role of functional traits, defined as morphological, physiological, or phenological characteristics measurable at the individual level that directly influence plant adaptation, survival, growth, and reproduction, and thus determine their functioning [112]. These traits are essential for understanding how different species contribute to the provision and long-term maintenance of ecosystem services [92, 113]. In this context, urban forestry should prioritize the enhancement of tree longevity and health, given that mature, vigorous individuals provide disproportionately greater benefits to urban environments [114]. Large-diameter and long-lived trees constitute critical components of urban green infrastructure due to their unique ecological contributions and structural stability [115, 116]. With fuller crowns and more extensive leaf surface areas, these trees enhance ecosystem functioning more effectively than smaller or younger individuals [117–120]. The magnitude of these contributions can vary not only across species [113, 121, 122] but also within species, depending on age and size. Therefore, the conservation and management of urban forests must focus on supporting the survival and performance of such individuals to optimize ecosystem service delivery and strengthen urban resilience in the face of accelerating climate change [123].

Large trees, despite their disproportionately critical role in the delivery of ecosystem services, are subject to the highest number of threats and are experiencing a pronounced global decline [115]. In the face of accelerating climate change, urban forestry must move

beyond treating species as taxonomically fixed and homogeneous units. Instead, it should recognize and incorporate the significant genetic variation that exists within species. Urban planning and tree selection strategies must proactively identify and promote genotypes or ecotypes that are best suited to future climatic conditions, thereby ensuring long-term planting success and optimizing ecosystem service returns on investment [124]. The urgency of this adaptive approach is underscored by recent findings from Esperon-Rodriguez et al. [125], who reported that projected urban climate conditions have already exceeded the climatic tolerance thresholds of 56–65% of tree species currently planted across 164 cities in 75 countries. This alarming mismatch highlights the necessity of shifting toward the selection of climate-resilient species and provenances that are not only capable of surviving but also of thriving in increasingly stressful and unpredictable urban environments [126].

In this context, promoting intraspecific variation and focusing on functional traits is essential. Traits such as specific leaf area, photosynthetic rate, and water-use strategies determine how trees respond to urban stressors like drought, heat, and pollution, ultimately shaping their capacity to provide key ecosystem services [127–129]. Recent research underscores that functional trait diversity and plasticity enable urban trees to adapt to the heterogeneous conditions characteristic of cities, thereby enhancing resilience and ensuring sustained ecosystem functionality. Individuals originating from harsher climates or stressful growing conditions, environments analogous to projected future climates, may possess adaptive traits that increase their likelihood of survival [130, 131]. For example, sourcing genotypes from warmer parts of a species' distribution may introduce beneficial traits that improve adaptive capacity [132]. The success of urban greening under climate change hinges on phenotypic plasticity and the standing genetic variation available for natural or artificial selection [94, 133].

Age structure diversification is also a critical consideration in urban forestry. Urban forests, particularly street tree populations, frequently exhibit limited age heterogeneity [134], which significantly increases their vulnerability to synchronous decline or failure. Such simultaneous losses can lead to abrupt and prolonged disruptions in the provision of essential ecosystem services. This vulnerability is further exacerbated by escalating pressures from climate change, pests, and diseases, underscoring the imperative to incorporate a broader range of age classes in urban tree populations to enhance their resilience and long-term sustainability.

Species diversification has generated significant debate regarding the use of native versus non-native species in urban greenery, particularly concerning the invasion potential of exotic species. According to a recent review by Kumar and Singh [135], invasive alien plant species alter native ecosystems, weakening essential ecosystem services and ultimately harming human health. Furthermore, their rapid expansion and resource competition displace native flora, exacerbate environmental pollution, and generate elevated socioeconomic costs, highlighting the urgency of transdisciplinary approaches for sustainable management. Silva et al. [136] observed that exotic tree species exhibit greater reproductive capacity than native species, and this high reproductive capacity represents a significant risk to native species by competing for pollinators, altering existing pollination networks, and potentially facilitating invasion processes that can displace native flora. They suggest that for tropical urban ecosystems, where an

extensive species pool exists, native species should constitute the first option.

In Europe, Sjöman et al. [137] argue that species selection cannot categorically exclude exotic tree species, especially in regions with limited native species catalogs that prove insufficient to satisfy the ecosystem services and resilience required in harsh urban environments, a consideration that becomes even more critical in the face of climate change [125, 138, 139]. “Native-only” approaches may compromise urban ecosystem resilience, particularly in regions with extreme environmental conditions where native species lack the necessary tolerance to survive in built environments. While legitimate invasion risks may exist with some non-native species that can escape cultivation and become invasive [140], they also highlight their potential inability to support native fauna diversity comparable to their native counterparts, resulting in decreased food web complexity [141–146].

Arguments favoring non-native species include their higher growth rates and proper development of fine root biomass systems under drought conditions when originating from geographically adapted climatic areas [147], high capacity for ecosystem service provision [140, 148], and potential to contribute to climate-resilient species assemblages when selected using ecological trait-based approaches, including isohydric behavior or characteristics such as leaf critical temperature, leaf water potential at turgor loss point, or stomatal conductance [124, 149–151]. These attributes could translate into more rapid, durable, and resilient ecosystem service provision.

Such arguments, both for and against, lead to the necessity of context-specific management, consistent with this manuscript's proposal, given the existing trade-off between invasion risks and provision-resilience benefits. Site-specific risk assessments that consider bioregional context and current and projected climatic conditions are therefore necessary, adapting urban policies on non-native tree use to specific surrounding environmental characteristics of the intervention area rather than adopting blanket prohibitions across the entire urban matrix.

A comprehensive understanding of urban tree performance must begin below ground, where many of the most critical yet frequently overlooked stressors are found. Root development and soil conditions fundamentally influence tree health, structural form, and longevity; however, these factors remain insufficiently integrated into urban forestry planning and practice [123]. Urban soils are often compacted, poorly aerated, and chemically imbalanced, creating inhospitable environments that severely constrain root growth and function [152]. These below-ground limitations compromise tree stability, water and nutrient uptake, and overall resilience, constraints that cannot be fully mitigated by above-ground interventions alone. Consequently, to secure the long-term viability and functionality of urban green infrastructure, prioritizing soil quality, adequate rooting space, and optimal subsurface conditions is essential in tree selection, planting, and maintenance protocols. Recent tree-planting strategies, such as structural cells, structural soils, and enhanced tree pit systems (particularly suspended pavement systems), have demonstrated significant positive outcomes. These include increased root colonization in the soil, higher growth rates, more extensive and better-developed root systems that support larger and healthier trees and improved mechanical stability when compared with conventional tree pit plantings [153–156].

While urban trees provide numerous ecosystem services, it is crucial to recognize their potential to generate disservices. These negative impacts include economic costs associated with maintenance, property damage, safety hazards arising from reduced visibility or increased concealment, and nuisances such as leaf litter and root intrusion [157]. Among the most studied ecological disservices are the emission of biogenic volatile organic compounds, which react with nitrogen oxides to form ground-level ozone, thereby exacerbating urban air pollution [158], and the production of allergenic pollen, which poses health risks for susceptible populations [159]. These challenges highlight the critical importance of informed species selection. Implementing site-appropriate species not only optimizes the delivery of ecosystem services but also minimizes disservices, ensuring that urban forestry initiatives promote both public health and sustainable urban environments.

Finally, beyond the imperative to avoid uncritical increases in tree quantity and diversity, spatial distribution must be recognized as a fundamental priority. The literature consistently shows that socioeconomically disadvantaged groups are both exposed to the most harmful environmental conditions and have the least access to the benefits provided by urban trees.

Multiple studies have documented the strong link between environmental inequality and health outcomes [160, 161]. Low-socioeconomic status communities face higher exposure to air pollutants and other environmental hazards [162–164]. Furthermore, research on how socioeconomic position influences exposure disparities in air pollution across ethnic groups has found that these inequities persist even after adjusting for socioeconomic status [165, 166]. In both the U.S. and the Netherlands, minority ethnic groups experience higher pollutant concentrations than the majority population [164, 167, 168]. Regarding the urban heat island, individuals with lower socioeconomic status and pre-existing health conditions endure greater exposure and vulnerability to its adverse effects [169, 170]. In U.S. cities, Black residents and those living below the poverty line experience higher urban heat intensities [171]. Aznárez et al. [172] revealed significant mismatches between cooling supply and demand in Vitoria-Gasteiz, where socioeconomically disadvantaged areas face pronounced deficits, whereas more affluent neighborhoods enjoy privileged access to urban vegetation's cooling benefits.

Regarding accessibility to vegetation and its services, research has consistently identified correlations between neighborhood income and vegetation greenness [173, 174]. These studies highlight the uneven distribution of urban canopy cover, resulting in inequitable access to ecosystem services, largely influenced by socioeconomic factors [169]. This socioeconomic disparity, often referred to as the “luxury hypothesis,” suggests that wealthier residents have greater financial resources to invest in the planting and maintenance of vegetated landscapes [175]. Numerous studies have documented that tree cover is frequently unevenly allocated among communities with diverse demographic and socioeconomic characteristics, with lower-income neighborhoods and ethnic minorities consistently experiencing significantly lower canopy coverage [35, 176–179]. Strategically directing planting efforts to promote equitable tree distribution addresses critical social and environmental justice concerns and advances progress toward SDG 11, which aims to create inclusive, safe, resilient, and sustainable cities.

Given that the discussion presented is highly focused on urban contexts and could lead to the incorrect conclusion that the authors suggest that the urban forest is composed solely of street trees and those existing in parks and gardens, it is also necessary to emphasize the central role that natural forests near the city, peri-urban forests, and even isolated forest patches existing within the urban matrix play in the provision of ecosystem services [180–182]. Although the manuscript and the vision presented suggest how new urban tree planting initiatives should be carried out, it should also be emphasized that integrative plans must constitute the foundation. Such documents should include strategies for new plantations, along with specific sections that follow best practices in the management of these more natural forest patches [183–189]. The urban forest is composed of the entire set of trees existing in the urban matrix or in its proximity directly accessible to the population, and following the line of what is proposed in the article, its management must be context-specific, understanding that urban trees and parks and gardens have different characteristics from forests and that our strategies must also contemplate these differences.

## 5. Conclusions

Can cities serve simultaneously as refuges and engines of biodiversity? The answer is yes, but only if urban planning adopts a multifaceted and strategic approach. This includes minimizing territorial fragmentation, integrating biomimicry principles into built environments, and establishing concentric green infrastructures that connect natural areas with urban interiors through corridors and nodal parks. Additionally, the use of exotic yet ecosystem service-efficient species can help cities adapt to heat island microclimates. Embracing advances in synthetic biology, under the safeguards of frameworks such as the Aichi Targets, can further facilitate the development of climate-adaptive urban species. Designing multi-layered urban forests and microforests to maximize functional diversity completes this vision, enhancing ecosystem resilience and service provision.

Should increasing plant biodiversity be the primary objective of urban greening in cities? Emphatically not. True urban sustainability demands a paradigm shift, from simply counting species to designing complex, adaptive, and equitable ecosystems. With thoughtful planning, cities can evolve into resilient, life-sustaining habitats that benefit both people and nature. Species diversification represents a step toward integrating the multiple ecological, social, and functional factors previously outlined into a unified framework. However, the critical question remains: how much diversity is truly necessary, and based on what species composition? Simply increasing diversity without considering the full range of relevant factors risks falling into the same flawed randomness Jaynes criticized, which leads to ineffective and fragmented urban forestry. We argue that a holistic approach is essential, as it determines which species to select, the functional and structural traits to prioritize, and where to plant them, ensuring the continuous provision of ecosystem services. In this context, diversity, or biodiversity, becomes a secondary outcome of a scientifically grounded selection process designed to maximize survival and sustained ecosystem function. The expected result is diverse urban forests, but the answer to “how much diversity is needed” ultimately depends on the factors that govern species survival and the ongoing delivery of ecosystem services.

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## Author contributions

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