

1 **Analyzing environmental factors that favor the growth of the invasive brown**
2 **macroalga *Rugulopteryx okamurae* (Ochrophyta): The probable role of the**
3 **nutrient excess**

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6 Jesús M. Mercado¹, Francisco Gómez-Jakobsen¹, Nathalie Korbee², Antonio Aviles²,
7 José Bonomi-Barufi³, María Muñoz⁴, Andreas Reul², Félix L. Figueroa²

8

9 ¹Centro Oceanográfico de Málaga, Instituto Español de Oceanografía-CSIC, Puerto
10 Pesquero s/n. 29640, Fuengirola, Málaga. Spain

11 ²Universidad de Málaga, Instituto de Biotecnología y Desarrollo Azul (IBYDA).
12 Departamento de Ecología y Geología. Facultad de Ciencias. Campus Universitario de
13 Teatinos s/n 29071 Málaga, Spain

14 ³Phycology Laboratory, Botany Department, Federal University of Santa Catarina,
15 88049-900, Florianopolis, SC, Brazil

16 ⁴Centro Oceanográfico de Baleares, Instituto Español de Oceanografía-CSIC, Muelle de
17 Poniente s/n, 07015, Palma de Mallorca, Spain

18

19 *Corresponding author: jesus.mercado@ieo.es

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21 **ABSTRACT**

22 Time series of temperature, salinity and nutrients in the Strait of Gibraltar (SoG) were
23 researched to analyze which factors explain the invasive success of *Rugulopteryx*
24 *okamurae*, which has colonized wide coastal areas at the Spanish and Moroccan coasts
25 since 2016. Temperature and salinity were higher in the SoG compared to its native
26 habitat, implying that the alga is active during the whole seasonal cycle and grows
27 optimally at the high salinities occurring in the SoG. Nitrate removal experiments
28 indicate that the alga is able to linearly increase its N uptake rates following boost in
29 nitrate concentration. Furthermore, *R. okamurae* N content ranged from 1.4% to 4.5%
30 suggesting that this species has high N storage capacity potentially usable when the
31 external N concentration decreases. These physiological characteristics would explain
32 sharp growth of the alga in the SoG where high N concentrations are registered
33 occasionally.

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36 **Key words:** inorganic nutrients, invasive species, nitrate uptake, *Rugulopteryx*
37 *okamurae*, salinity, temperature.

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40 **1. Introduction**

41 Regional studies have identified hundreds of non-indigenous marine species introduced
42 by humans, including tens of macroalgae (Ribera Siguan, 2002; Zenetos and Galanili,
43 2020). At least 30% of non-indigenous macroalgae have produced impacts on the native
44 ecosystems worldwide (Davidson et al., 2015) such as monopolization of the space and
45 drastic reduction of the abundance and diversity of the native species, which are
46 indicative of invasive behavior. Little attention has been paid to the reasons why some
47 macroalgal species which do not feature an invasive behavior in their natural
48 distribution areas, are able to colonize opportunistically other areas once entered.
49 Overall, ecosystem imbalances produced by natural and/or anthropogenic changes
50 would explain the success of introduced species (Shaffelke et al., 2006). Notorious
51 examples of seaweed invasions or extreme algal blooms favored by human
52 perturbations are those ones produced by *Undaria pinnatifida* in coastal areas of Europe
53 and Argentina (Meretta et al., 2012; Kraan, 2016), *Ulva prolifera* in the Yellow sea and
54 East China Sea (Zhao et al., 2019), *Sargassum* sp. in the Caribbean sea (Franks et al.,
55 2016; Wang et al., 2019; Chávez et al., 2020) and *Caulerpa taxifolia* in the
56 Mediterranean Sea (Meinesz et al., 1993, 2001).

57 It is widely accepted that the climate change combined with other
58 anthropogenic impacts (e.g. eutrophication and acidification) play a major role in
59 determining the colonizing success of non-native species (Diez et al., 2012). In fact,
60 anthropogenic disturbances have been identified as the further determining factor for
61 some seaweeds invasions (Davidson et al., 2015). For instance, physical alterations of
62 habitat produced by storms (Scheibling and Gagnon, 2006) and/or great amount of
63 nutrient discharges associated to pollution gave competitive advantage to some
64 opportunistic non-native species in different ecosystems (Piazzi et al. 2001, 2005, 2016;

65 Boudouresque and Berlaque, 2002; Piazzi and Balata 2009; Lapointe and Bedford,
66 2011; Gennaro and Piazzi, 2014). Biotic factors such as release of natural herbivorous
67 with feeding preference on the native communities might also contribute to the success
68 of the invasive macroalgae (Vermeij et al., 2009, Noè et al., 2018).

69 The Mediterranean Sea is the region with the highest number of introduced
70 macroalgal species with more than hundred non-indigenous species documented so far
71 (Boudouresque and Verlaque, 2005; Williams and Smith, 2007). Approximately a
72 dozen of these species have shown invasive traits (Verlaque and Fritayre, 1994; Piazzi
73 and Cinelli, 2003). The brown macroalga *Rugulopteryx okamurae* (Dawson)
74 I.K.Hwang, W.J. Lee & H.S.Kim (Dictyotales, Ochrophyta) has been recently added to
75 this list of invasive species due to it is invading the coasts of the Alboran Sea in the
76 Western Mediterranean Sea from 2016 (García-Gómez et al., 2017, 2020; El Aamri et
77 al., 2018). Recently it is starting to invade areas in Provence coast in France (Ruitton et
78 al., 2021) and the Portugal coast (Faria et al., 2021). This algal species originated from
79 East Asia (China, Japan, Korea; DeClerck et al., 2006) was identified by the first time in
80 the Mediterranean Sea by 2002 in the Thau Lagoon, where its entering was associated
81 to oyster aquaculture (Verlaque et al., 2009). The alga grew in the lagoon from that date
82 (Boudouresque et al., 2011) without showing an invasive trait. In contrast, from its first
83 detection in 2015, *R. okamurae* has colonized wide areas of the rocky shore in the
84 southern and northern coasts of the Strait of Gibraltar (SoG). Large biomass
85 accumulations were recorded for the first time in July 2016 (Altamirano-Jeschke et al.,
86 2016) reaching its maximum growth in 2017 when almost 100% of coastal habitats
87 from surface to 15 m depth at the SoG north part (within the marine protected area
88 *Natural Park of The Estrecho*) were occupied. Consequently, only two years from its
89 first detection, drastic alterations in the biodiversity of the native communities

90 coinciding with sharp growth of *R. okamurae* have been produced (García-Gómez et al.,
91 2021). Nowadays, its impact continues to be severe since thalli of this species are
92 frequently washed ashore in the north of the Alboran Sea, forcing their cleaning and
93 impacting significantly on the tourism activity. The nets of fishing vessels in the SoG
94 are often collapsed by *R. okamurae* floating thalli even in areas far from the shore,
95 producing significant economic loss.

96 Despite the impacts produced by *R. okamurae*, the factors that triggered this
97 invasion in the SoG still remain little known. Most studies published so far mention
98 simplistically the temperature as a possible cause favoring the invasion. However, the
99 oceanic circulation at the SoG produces conspicuous interannual, seasonal and short-
100 term variability in temperature, salinity and nutrients (García-LaFuente et al., 2000;
101 Gómez et al., 2004; Gómez-Jakobsen et al., 2019); additionally, the area receives
102 significant inputs of inorganic nutrients from land sources (Mercado et al., 2018).
103 Therefore, it is tricky to attribute the invasion of *R. okamurae* to a particular
104 environmental factor without an exhaustive analysis of its variability patterns. In this
105 work, we analyzed time series of temperature, salinity and inorganic nutrients covering
106 a wide time period that includes the first phases of the invasion. The environmental
107 conditions in the study area were compared with the origin habitat of *R. okamurae* and its
108 possible competitive advantages were researched. Particularly, we investigated if the
109 alga capacity for assimilating dissolved inorganic nitrogen could be relevant in
110 determining its success in the SoG. Based on these analyses, we hypothesized that
111 changes in the hydrological circulation patterns combined with nutrient pollution might
112 be an important factor that favored the growth of *R. okamurae* at the study area.

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115 **2. Material and Methods**

116

117 *2.1. Gathering of meteorological and oceanography data*

118 Time series of air and water temperature and satellite chlorophyll *a* were built
119 with daily data gathered for the period 2010-2019. The daily data of air temperature
120 were obtained from a meteorological station located in the town of Tarifa (Spanish
121 Meteorological Agency, <http://www.aemet.es>; Fig. 1). Daily means of temperature
122 provided by a temperature sensor moored in the coast at 33 m depth (36.00°N, 5.59°W;
123 Puertos del Estado) were also collected for the same period. Ocean color data for a
124 location nearby the Strait of Gibraltar were retrieved from the platform MODIS-Aqua
125 (reflectances at 443, 488 and 547 nm were used). This satellite provides a daily image
126 of the study zone with a spatial resolution of 1.1 km². Level 2 scenes (reprocessing
127 2018) from August 2002 to October 2019 were downloaded from the NASA-Ocean
128 Color website (<https://oceancolor.gsfc.nasa.gov/>). Those scenes inadequate for the
129 analysis due to sun glitter and/or the presence of clouds or fog were discarded, and only
130 valid scenes were used. Satellite chlorophyll *a* concentrations were calculated from the
131 reflectance values by using the algorithm SMED3M (Gómez-Jakobsen et al., 2018).
132 The daily data were weekly averaged prior to be analyzed statistically.

133 Outputs of a 3D high-resolution biogeochemical multi-year model for the SoG
134 (36.00°N, 5.58°W; Fig. 1) were downloaded from the E.U. Copernicus Marine Service
135 (<https://resources.marine.copernicus.eu>; product: IBI_MULTIYEAR_BGC_005_003;
136 processing level: L4). IBI-MFC product is based in an application of the
137 biogeochemical model PISCES running simultaneously with the ocean physical IBI
138 reanalysis, generating both products at the same horizontal resolution of 1/12°. The
139 retrieved modeled variables at 1 m depth included salinity, nitrate, phosphate, mixed

140 layer depth (MLD), transparency and meridional (S-N) and zonal (W-E) components of
141 the current velocity. Monthly time series of each variable were used to analyze trends.

142 In order to research the nutrient conditions in the invaded area, concentrations of
143 nitrate, nitrite, ammonium and phosphate obtained in coastal stations located nearby or
144 within the invaded area were downloaded from a database of the Andalusian Regional
145 Government (*Visor de la Calidad de las Aguas de Andalucía*
146 https://laboratorioediam.cica.es/Visor_DMA; Consejería de Agricultura, Ganadería,
147 Pesca y Desarrollo Sostenible de la Junta de Andalucía; Fig. 1). These stations are
148 sampled several times each year within monitoring programs for implementation of the
149 EU Water Framework Directive. Additionally, nutrient concentrations in the water
150 column at a station located in the outer part of the Algeciras Bay (36.02°N, -5.38°W;
151 Fig. 1) were used to characterize the nutrient composition of the seawater that flows in
152 the study area. These offshore station data were obtained during different research
153 surveys performed quarterly in 2011-2012, 2014-2015 and 2019 by the Spanish Institute
154 of Oceanography-CSIC. In these cruises, water samples were obtained at different
155 depths with Niskin bottles and the nutrients were analyzed in the laboratory following
156 the protocols published in Ramírez et al. (2005) by means of segmented flow analyzer
157 QuAAtro 39 (Seal Analytical).

158

159 *2.2. Measurements of the physiological performance*

160 Experiments of nitrate removal were conducted with algae collected at the Caleta
161 beach (Tarifa) on 16th October 2019. Thalli attached to rocks were selected and rapidly
162 transported in icebox to the laboratory where were washed with seawater and epiphytes
163 were removed. About 2 g of cleaned thalli were placed into 1 L methacrylate flasks
164 containing 38.0 salinity seawater enriched with 50 μM of NaNO_3 as N source at 22 °C.

165 The culture was vigorously aerated and illuminated with 200 $\mu\text{mol photons m}^{-2} \text{s}^{-1}$ of
166 irradiance emitted from white light LEDs (RGB) at photoperiod of 12h:12h (light:dark).
167 Thalli acclimated to these conditions for one week were used to determine the nitrate
168 uptake kinetics. Different initial concentrations of nitrate were tested by adding small
169 amounts of a solution of NaNO_3 to the culture flasks containing fresh weight (FW) 2 g
170 of acclimated thalli. The nominal initial concentrations of nitrate were: 50, 100, 160,
171 250, 400 and 800 μM . Each concentration was tested by triplicate for 7 days. Seawater
172 samples were collected at the beginning of the experiment and after 1, 5 and 7 days.
173 Nitrate concentrations were estimated by means of segmented flow analyzer QuAAtro
174 39 (Seal Analytical). Daily nitrate removal rates (NRR) were estimated in each
175 incubation flask from the disappearance of nitrate after the first day of treatment. The
176 NRR vs. initial nitrate concentration curve was fitted to the Michaelis-Menten model by
177 non-linear least squares fitting. Maximal NRR (NRR_m) and semi-saturation
178 concentration (K_m) were estimated from the fitting parameters.

179 The carbon and nitrogen contents of thallus pieces collected in the beach of Caleta
180 (Tarifa) in February, March, June, July and August 2020 were determined. The total
181 carbon and nitrogen contents of crude extracts were quantified using an elemental
182 analyzer bounded to an IR detector CNHS LECO-932 (LECO Corp. MI, USA) by
183 burning 10 mg of dry samples at 1000°C.

184

185 2.3. Statistical analyses

186 The weekly time series of air and water temperature and satellite chlorophyll *a* as well
187 as the monthly time series of the modeled variables were analyzed by using the
188 technique of seasonal decomposition. The seasonal, trend and irregular components of
189 the time series were calculated with the function *Decompose* in R that performs a

190 classical seasonal decomposition by moving averages (Kendall and Stuart, 1983).
191 Additionally, the daily time series of water temperature was used to analyze the
192 frequency and intensity of extreme temperature events. The R package *heatwaveR*
193 specifically designed to detect heatwaves and cold spells was used (Schlegel and Smit,
194 2018). The functions *ts2clm* and *detect_event* were used to detect duration and intensity
195 of heatwave events defined as periods longer than four days for which the temperature
196 was higher than the seasonally varying threshold value (90th percentile threshold;
197 Hobday et al. 2016). The function *mcs* was used to detect cold spells defined as periods
198 longer than four days for which temperature was lower than the seasonally varying 10th
199 percentile threshold.

200 In order to research the changes in the hydro-geochemical patterns that occurred
201 during the invasion, a principal component analysis (PCA) was performed with the
202 monthly trend series of the variables obtained from the biogeochemical model of the
203 E.U. Copernicus Marine Service. PCA permits identifying the main modes of variation
204 of the hydro-geochemical variables as well as different time periods characterized by
205 shifts in one or various variables. PCA run in R using the function *rda* of the package
206 *Vegan* (Oksanen et al., 2014). The variables included in the analysis were temperature,
207 salinity, transparency, nitrate, phosphate, chlorophyll *a*, mixed layer depth and velocity
208 zonal and meridional components of the surface current. The variables were
209 standardized prior to the analysis.

210 The differences in nutrient concentrations and salinity between coastal stations
211 and the offshore station were tested with a Kluskal-Wallis test at $p < 0.05$.

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213

214

215 **3. Results**

216

217 *3.1. Hydrology and nutrients*

218 The weekly averaged time series showed maximal temperatures in air (26.5°C)
219 and water (22.8°C) in July 2017 and July 2016, respectively. Concordantly, the
220 temperature trend time series indicate that there was a warming period from 2015 to
221 2018 in both air and water (Fig. 2). Note that the intensity of this warming was
222 smoother in water compared to air. Interestingly, the most extreme values of weekly
223 temperature in water were obtained in winter and summer of 2016. The analysis of
224 temperature extreme events based on the daily temperature time series indicates that
225 there were 17 heatwave events in 2010-2020, 58% of which were produced during
226 2015-2019 (Suppl. Fig. 1) although the amount of days with heatwaves in 2015-2019
227 and 2010-2014 was similar. The longest heatwave was produced in 2010 while the most
228 intensive heatwave was registered in 2016. The occurrence frequencies of cold spells in
229 2010-2014 and 2015-2019 were also similar. The most intensive cold spell occurred in
230 2015.

231 Satellite chlorophyll *a* in the study area averaged 0.3 $\mu\text{g L}^{-1}$, with annual peaks
232 above 0.6 $\mu\text{g L}^{-1}$ that were normally obtained in summer (Suppl. Fig. 2). However, it is
233 remarkable that in 2015 and 2016, annual peaks of satellite chlorophyll *a* were not
234 registered; in fact, the lowest weekly concentrations of chlorophyll *a* were obtained
235 during these two annual periods when there was a decreasing trend compared to the rest
236 of the time series.

237 The time series of the modeled variables indicate that hydrodynamic changed in
238 the study area in 2015-2016 as surface salinity decreased and velocity zonal component
239 increased in comparison to previous periods (Fig. 3; note that positive values of this

240 component indicate that the flow runs from west to east). The highest positive value in
241 the irregular component time series of the zonal current velocity for the whole study
242 period was produced in 2015. The outputs of the model also point out that surface
243 nutrient concentrations increased in 2015, with annual maxima in 2016-2018 that were
244 notably higher than the maxima modeled for previous years (Fig. 4).

245 The results of the PCA performed with the trend time series of the modeled
246 variables also illustrate clearly these inter-annual changes. The first PC extracted (PC1)
247 explained 43% of the variability (Fig. 5a). Transparency and meridional component of
248 the current velocity were the most important variables that contributed positively to PC1
249 while nutrients and chlorophyll *a* did negatively. Therefore, this PC was related to
250 changes in phytoplankton biomass which were associated to shifts in nutrient
251 concentration and displacement velocity of the surface water from the north coast to the
252 south. Salinity and current velocity zonal component were the main contributors to PC2
253 (20%) although the signs of the contributions were different. It can be assessed that this
254 PC indicates changes in the horizontal transport velocity of the surface water masses
255 that circulate throughout the Strait of Gibraltar. Note that temperature did not contribute
256 to these two first variance components; however, this variable was the main contributor
257 to PC3 (13% of explained variability) which indicates that part of the variability in the
258 surface temperature trend series was not associated to hydrological variability. PC1
259 scores decreased progressively from 2010 to 2019 (Fig. 5b), which can be interpreted as
260 higher nutrient concentration was associated to meridional current velocity decreasing.
261 The variation time patterns of PC2 differed compared to PC1 given that the most
262 negative scores were obtained in the period 2013-2017 indicating that the zonal current
263 velocity increased during that period. It has to be noted that salinity and MLD were
264 positively correlated to PC2 (Fig. 5c), therefore during that period of zonal current

265 velocity increasing, salinity and MLD reduced. Time series of the PC3 scores indicate
266 that only a few monthly periods (roughly distributed every 3 years) contributed
267 significantly to this PC from 2010-2014. However, it is notable that distribution of PC3
268 scores was more irregular since 2015, with some monthly periods presenting negative
269 scores.

270 The variability in the concentrations of dissolved inorganic nitrogen (DIN)
271 was higher in the coastal stations than in the offshore station (Fig. 6; note that there are
272 not ammonium data available for open sea). Nitrate concentration averaged in the
273 coastal stations ranged from 1.0 to 3.5 μM with the exception of 2017 when the average
274 was lower than 0.7 μM . These yearly concentrations were higher than in the 20 m
275 surface layer of the offshore station (apart of 2017) and similar to the concentrations in
276 the offshore deeper layer. However, the nitrate maxima in the coast were notably higher
277 than in the offshore deeper layer (Table 1) as also occurred with nitrite. Ammonium
278 concentrations were fairly high in comparison to nitrate, especially in 2017 when the
279 concentration maximum reached 20 μM . Phosphate concentrations at the coastal
280 stations were less unstable than nitrate and similar to offshore. These data indicate that
281 the coastal stations were enriched in DIN relative to phosphate in comparison to
282 offshore. The nutrient concentrations varied irrespective of the salinity in the coastal
283 stations (Suppl Fig. 3) where lower salinity was obtained in 2014 and 2015, matching
284 the decreasing trend in modeled salinity (Fig. 3). In comparison to the offshore surface
285 layer, the coastal stations had lower salinity.

286

287 3.2. *Physiological features*

288 The results of the nitrate removal experiments indicate that *R. okamurae* has
289 high capacity for using dissolved inorganic nitrogen (Fig. 7a) as nitrate was almost

290 totally depleted after only one day of culture in the treatments with initial concentrations
291 of 50 and 120 μM . In the treatments with the highest nitrate concentrations, nitrate
292 removal followed a linear kinetic indicating that NRR kept roughly constant during the
293 incubations. NRR calculated from removal after 24 h incubation followed saturation
294 kinetics reaching the maximal rate at the highest concentration assayed (Fig. 7b). The
295 semisaturation concentration (K_m) calculated by fitting these data to the Michaelis-
296 Menten model was 135 μM and NRR at nitrate saturation (NRR_m) was 83 $\mu\text{mol N}$
297 $\text{gFW}^{-1} \text{d}^{-1}$.

298 Carbon content of *R. okamurae* thalli collected in 2020 ranged from 34.2% in
299 February to 42.1% in August. The lowest and highest N contents were obtained in June
300 (1.4%) and August (4.5%), respectively. Consequently, the N content and C:N molar
301 ratio in the algal tissue varied by about three-folds.

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303

304 **4. Discussion**

305

306 *4.1. Temperature, salinity and nutrient conditions in the Strait of Gibraltar*

307 The presence of *Rugulopteryx okamurae* in macroalgal communities of the
308 warm temperate western Pacific Ocean has been documented frequently (Hwang et al.,
309 2009; Kitayama and Lin, 2012; Lim et al., 2017; Jung and Choi 2020; Jung et al., 2020).
310 According to the published information, the alga grows in the shallow subtidal and
311 above 15 m depth. In its native habitats, *R. okamurae* rarely becomes dominant; in fact,
312 the maximal abundance reported in the literature is ca. 50 g FW m^{-2} and the highest
313 coverage percentage is less than 5% (Jung and Choi, 2020; Jung et al., 2020; Lim et al.,
314 2017). Consequently, the biomass and coverage reached by the alga in the SoG ($11 \cdot 10^3$

315 gFW m⁻² and almost 100% of coverage in some points of Tarifa; García-Gómez et al.,
316 2021) are indicative of growth rates not described before for this species in its natural
317 habitat. Obviously, this sharp growth only can be attributable to features of the recipient
318 environment which favor *R. okamurae* and/or are a detrimental to the native
319 communities (Davidson et al., 2015). Increasing temperature is one of the
320 environmental factors which would enhance the growth and competitive output of
321 macroalgal invaders (Cebrián and Ballesteros, 2010; Martínez-Luscher and Holmer,
322 2010). Comparisons between the SoG and native habitats of *R. okamurae* (Table 1)
323 indicate that, on average, maximal and mean temperatures in both areas are similar but
324 the temperature winter minimum is higher in the SoG (14.0 °C compared to 9.2 °C; Jung
325 and Choi, 2020; Jung et al., 2020; Lim et al., 2017). In its native habitat, *R. okamuare* is
326 present round-year although its maximal growth rate is reached in spring-summer (Jung
327 et al., 2020). The annual cycle of growth appears to be similar in the SoG (García-
328 Gómez et al., 2021); however, warmer winters will imply that the active growth annual
329 period of *R. okamurae* might be longer than in its native habitat (assuming that the
330 temperature is a factor that regulates its annual cycle). In fact, significant growth of the
331 alga has been described in Tarifa (northern SoG) even in winter (García-Gómez et al.,
332 2018). Salinity also differs between the two areas; in the western Pacific Ocean, it varies
333 within a reduced range below 34.5 while in the SoG salinity presents a wider variation
334 range with minima above 34.9 and maxima frequently above 38.0 (Table 1). The sharp
335 growth of *R. okamurae* in the SoG suggests that this species is euryhaline.

336 Nutrient regime in the SoG is characterized by presenting high variability
337 linked to modifications in speed (from 0.05 ms⁻¹ to 0.35 ms⁻¹) and/or entry angle of the
338 strong and permanent eastwards current of the surface Atlantic Water (the so-called
339 Atlantic Jet) that enters into the Alboran Sea (Sarhan et al., 2000; Sammartino et al.,

2015; Macias et al., 2016). Both, direction and strength of the flow (Sánchez-Garrido et al., 2013; Gómez-Jakobsen, 2019), influence significantly the surface circulation patterns, regulate the nutrient distribution and affect the productivity in the west side of the SoG (Sarhan et al., 2000, Reul et al. 2002, 2008). Furthermore, the Ekman transport favors upwellings of enriched waters in the own SoG (Macías et al., 2007), which are normally produced in Spring, when the phytoplankton annual bloom occurs (Ramírez et al., 2005). The highest nitrate concentrations in the surface waters are obtained under a certain combination of upwelling and tide cycle (Echevarría et al., 2002), contributing to increases in the N:Si:P ratio eastwards. However, the high DIN concentrations obtained in the coastal stations in our report (punctually higher than 20 μM) cannot be explained by hydrological variability. Furthermore, the phosphate concentration at these stations was low in comparison to DIN and the ratio DIN:P departed significantly from the molar ratio reported for SoG waters by Huertas et al. (2012). These results indicate that the study area is affected by important entries of nutrients from terrestrial sources as Mercado et al. (2018) suggested previously for the Bay of Algeciras based on data from 2012 to 2015. In the same sense, high concentrations of dissolved organic carbon produced by phytoplankton have been attributed to inputs from the rivers that flow in the west side of SoG, mainly the Guadalquivir river that receives enormous amount of agricultural leachate (Álvarez-Salgado et al., 2020). The increase in the zonal component of the velocity and reduction in the meridional component in 2016-2017 will imply that the inputs of terrestrial water from the Spanish coast will circulate closer to the northern coast. This change in the circulation pattern would contribute to local enrichment of the surface water in the coastal stations creating suitable conditions to stimulate the growth of *R. okamurae*.

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366 4.2. Contribution of the physiological performance of *R. okamurae* to the invasion

367 High nutrient concentrations are prerequisite for macroalgal blooms as the bloom-
368 forming species have nutrient assimilation elevated capacity compared to other algae
369 (Wang et al., 2020). According to its kinetics of nitrate removal, irregular supply of
370 high nutrients would give a competitive advantage to *Rugulopteryx okamurae* on native
371 macroalgae. The high NRR_m is comparable with uptake rates reported for other bloom-
372 forming macroalgae (Table 2) and implies that the alga will assimilate large amount of
373 nitrate when the external concentrations are elevated. Nitrate affinity of *R. okamurae* is
374 comparatively low taking into account values published for some of these bloom
375 forming macroalgae (from 10 to 61 μM ; see the data compiled in Wang et al. 2014).
376 However, the relatively high K_m obtained in our experiments indicates that *R. okamurae*
377 will respond to nitrate increases by linearly increasing its assimilation rates.
378 Furthermore, this species will keep constants its nitrate removal rates for several days of
379 growth under high nitrate unlike many other macroalgae that down-regulate the nitrate
380 assimilation mechanisms (e.g. see den Hann et al., 2016 and Wang et al., 2014).

381 The growth responses to N supply depend on the interaction with the internal N
382 pools; in fact, a relationship between ambient growth rates of *Ulva* spp. and thallus N
383 content has been described (Fong et al., 2001; Teichberg et al., 2007). N content of *R.*
384 *okamurae* was highly variable which would be indicative of storage capacity of N when
385 its external supply is increased, which is also a common feature among bloom forming
386 algae. For instance, it has been shown that the N internal pool of *U. lactuca* is linked to
387 external DIN supply and can vary between 1.5% and 5.8% (Fujita, 1985; Pedersen and
388 Borum, 1996; Teichberg et al., 2010). Furthermore, internally stored N would be used
389 by the alga for keeping high growth rates when the external concentrations decrease

390 (Lartigue and Sherman, 2006; Kennison et al., 2011). The fact of C content in *R.*
391 *okamurae* is high regardless of its N content, would be indicative of high carbon
392 fixation rates are reached even under low DIN concentrations, which is expected if the
393 alga uses internal N. Compared to macroalgae inhabiting rocky areas of the southern
394 Iberian Peninsula (collected close to the study area; see Mercado et al., 2009, for a more
395 detailed description), *R. okamurae* presented the highest C content (Suppl. Fig. 4). It is
396 also notable that the variation range of N content covered the whole range reported for
397 the macroalgae with N content higher than 1% collected at the southern Iberian
398 Peninsula. This variability in N content is consistent with the great morphological
399 plasticity that has been described for *R. okamurae* not only in the SoG (Figueroa et al.,
400 2020; Salido and Altamirano, 2020; García-Gómez et al., 2021) but also in its origin
401 place (Hwang et al., 2009). Note that physiological plasticity is a common characteristic
402 of the invasive algae (Davidson et al., 2015; Zanolla et al., 2019) which permits their
403 opportunistic growth when the external conditions are favorable.

404 It cannot be discarded that the changing hydrological conditions described in our
405 report also affected negatively the native communities. Thus, it is notable that 2016 was
406 the annual period with the most extreme temperatures (including the occurrence of an
407 intensive waveheat event) and that the winters of 2017-2018 were particularly warmer.
408 Furthermore, modeled salinity for this period decreased notably. It is possible that these
409 environmental changes decreased the competitive performance of the native
410 communities. Furthermore, the hydrological changes that occurred before the massive
411 presence of *R. okamurae* in 2016 could reduce the resilience capacity of the ecosystem
412 although damages on native communities were not registered prior the invasion (García-
413 Gómez et al., 2021).

414

415

416 In summary, our data indicate that the success of *R. okamurae* in the SoG could
417 be due to a combination of both its particular physiological characteristics and the
418 favorable environmental conditions for its growth, including the high DIN availability
419 possibly being discharged from terrestrial sources. At any rate, more research on the
420 eco-physiology of the alga is necessary to assess its invasive capacity which would be
421 not only based on its response to seasonal and short-term changes in temperature,
422 salinity and nutrients but also on biotic factors. For instance, recently Casal-Porras et al.
423 (2021) have shown that *R. okamurae* presents high dilkamural content , a toxic
424 compound for native herbivores. The surface circulation associated to the Atlantic jet
425 entering into the Alboran Sea (García Lafuente et al., 2000; Gómez-Jakobsen, 2019)
426 might favor the fast expansion and colonization of nearby coastal areas. Additionally,
427 the fishery activity is elevated in the Spanish coast, particularly the bottom trawling
428 fishing (Muñoz et al., 2018) which would pluck *R. okumare* and contribute to its spread
429 along the coast. Therefore, this species should be closely monitored in the southern
430 Europe.

431

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445 **Mercado JM:** conceptualization, writing, original draft, data collection and analysis;
446 **Gómez-Jakobsen F:** satellite and meteorological data collection, time series analysis,
447 review; **Avilés A:** analysis of nutrients, nutrient data collection, review and editing;
448 **Bonomi-Barufi J:** sample collection, design and performing of lab experiments, review
449 and editing; **Muñoz M:** conceptualization, modelled data collection, review and editing;
450 **Reul A:** conceptualization, modelled data collection, review and editing; **Korbee N:**
451 sample collection, design and performing of lab experiments, review and editing;
452 **Figuroa FL:** conceptualization, writing, review and editing, sample collection and lab
453 analysis.

454

455

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767 **Table 1.** Means of temperature and nutrient concentrations measured in coastal stations
 768 close to Tarifa for different time periods within 2010-2019. Maximum and minimum
 769 values are also shown. 'Natural habitats' indicate the values of temperature and salinity
 770 published in the literature for the growth native area of *Rugulopteryx okamurae*.

		2010-2014	2015-2016	2017-2019	Native habitats
T air (°C)	mean	17.5	18.8	18.0	-
	max	25.5	26.0	26.5	-
	min	9.3	10.9	9.4	-
T water (°C)	mean	17.1	17.4	17.2	17.1
	max	21.6	22.8	21.3	23.8
	min	14.1	14.0	14.8	9.2
Salinity	mean	36.34	36.57	36.03	-
	max	37.96	38.04	36.97	34.30
	min	34.92	35.33	35.26	33.15
Nitrate (µM)	mean	2.14	1.94	1.28	-
	median	1.81	1.50	1.14	-
	max	6.39	5.52	3.92	-
	SD	1.80	1.50	1.61	-
Nitrite (µM)	mean	0.40	0.29	0.27	-
	median	0.22	0.30	0.24	-
	max	4.13	0.72	0.93	-
	SD	0.66	0.05	0.06	-
Ammonium (µM)	mean	1.24	0.99	2.82	-
	median	0.35	0.87	1.30	-
	max	8.06	2.71	20.2	-
	SD	1.63	0.52	4.60	-
Dissolved inorganic nitrogen (µM)	mean	3.74	3.21	4.37	-
	median	2.88	2.79	2.73	-
	max	13.14	8.35	23.09	-
	SD	2.91	1.78	5.17	-
Phosphate (µM)	mean	0.08	0.12	0.08	-
	median	0.05	0.07	0.05	-
	max	0.45	0.58	0.20	-
	SD	0.06	0.13	0.05	-

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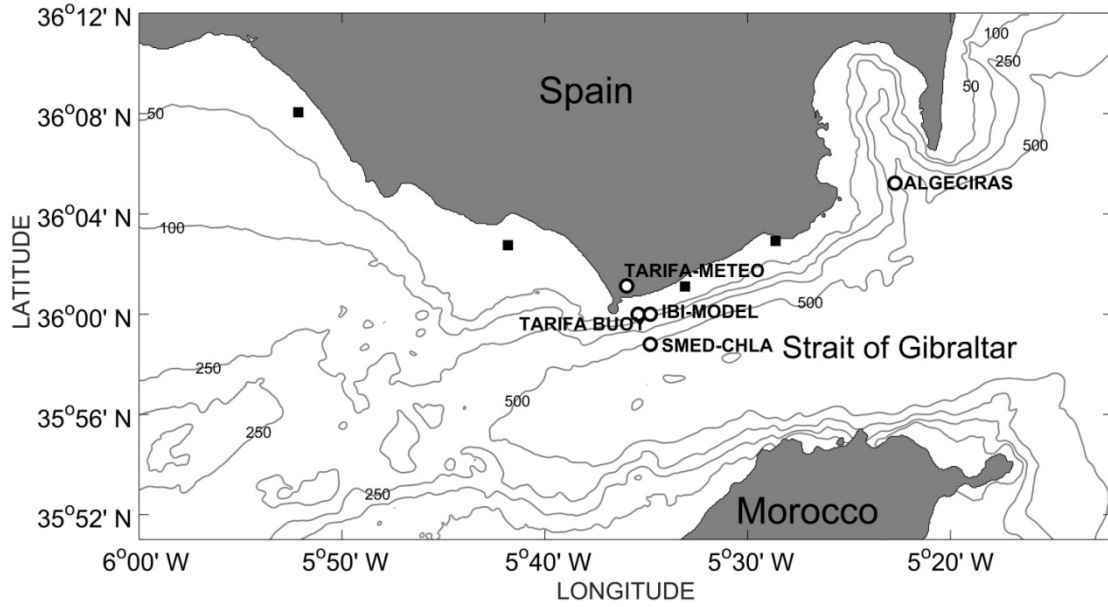
774 **Table 2.** Rates of nitrate removal (NRR) published for different macroalgae including
 775 some bloom-former species (only data expressed on fresh weight basis have been
 776 considered). The experimental conditions are indicated (T: temperature; I: $\mu\text{mol photons}$
 777 $\text{m}^{-2} \text{s}^{-1}$; P: photoperiod).

	NRR [$\mu\text{moles gFW}^{-1}\text{d}^{-1}$]	Experimental Conditions	Source
<i>R. okamurae</i>	83	T: 25°C; I: 250;P: 12:12	This work
<i>Porphyra leucosticta</i>	43	T: 15°C; I: 60; P: 12:12	Mercado et al. 1999
<i>Fucus vesiculosus</i>	10		
<i>Fucus serratus</i>	10	T: 15°C; I: 100;P: 10:14	Gordillo et al. 2002
<i>Laminaria digitata</i>	10		
<i>Ulva lactuca</i>	150	T: 25°C; I: 100; P: 12:12	Gordillo et al. 2001
<i>Achantaphora spicifera</i>	42.4	Outdoor conditions	Dailer et al. 2012
<i>Dictyota acutiloba</i>	50.6		
<i>Hypnea musciformis</i>	24.9		
<i>Ulva lactuca</i>	28.5		
<i>Caulerpa lentillifera</i>	150	T: 27°C; I: 150; P: 12:12	Bambaranda et al. (2019)

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780 **Fig. 1.** Position of the meteorological station (TARIFA-METEO), underwater probe of
781 temperature (TARIFA BUOY), sampling coastal stations where nutrient data were
782 obtained (solid black squares) and open sea station (ALGECIRAS). The coordinates of
783 the locations from which modeled (IBI-MODEL) and satellite data (SMED-CHLA)
784 were retrieved are also shown.
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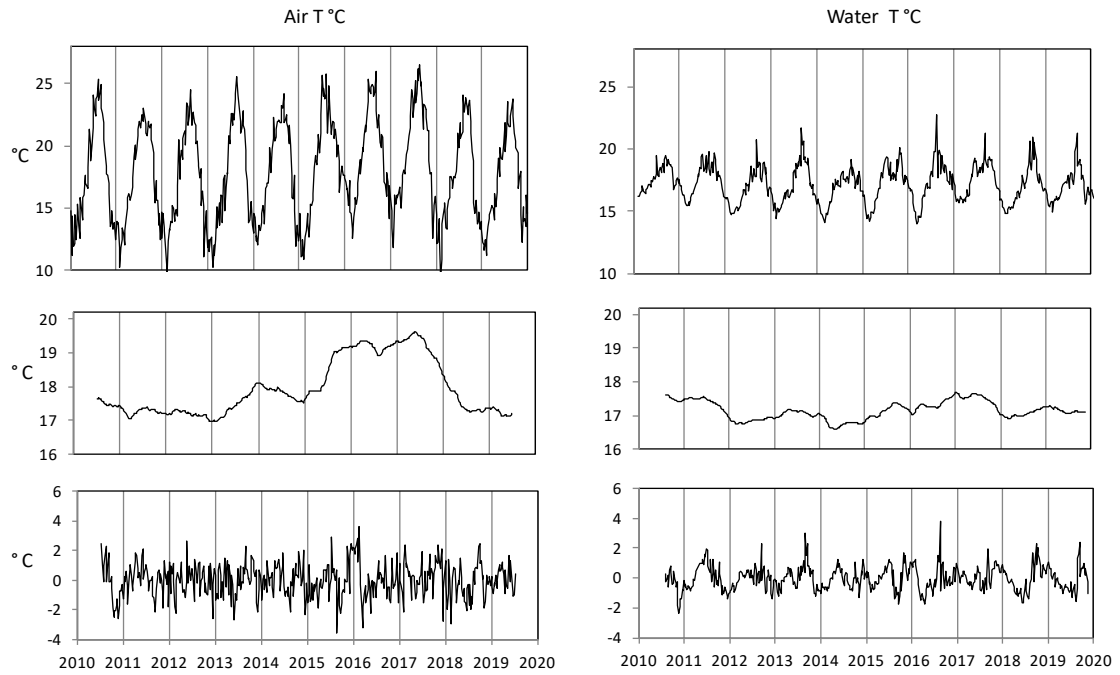
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791 **Fig. 2.** Time series of air and near surface water temperature. Upper panels show the
792 raw data series. Middle panels represents the trend series. Lower panels indicate the
793 irregular component time series.
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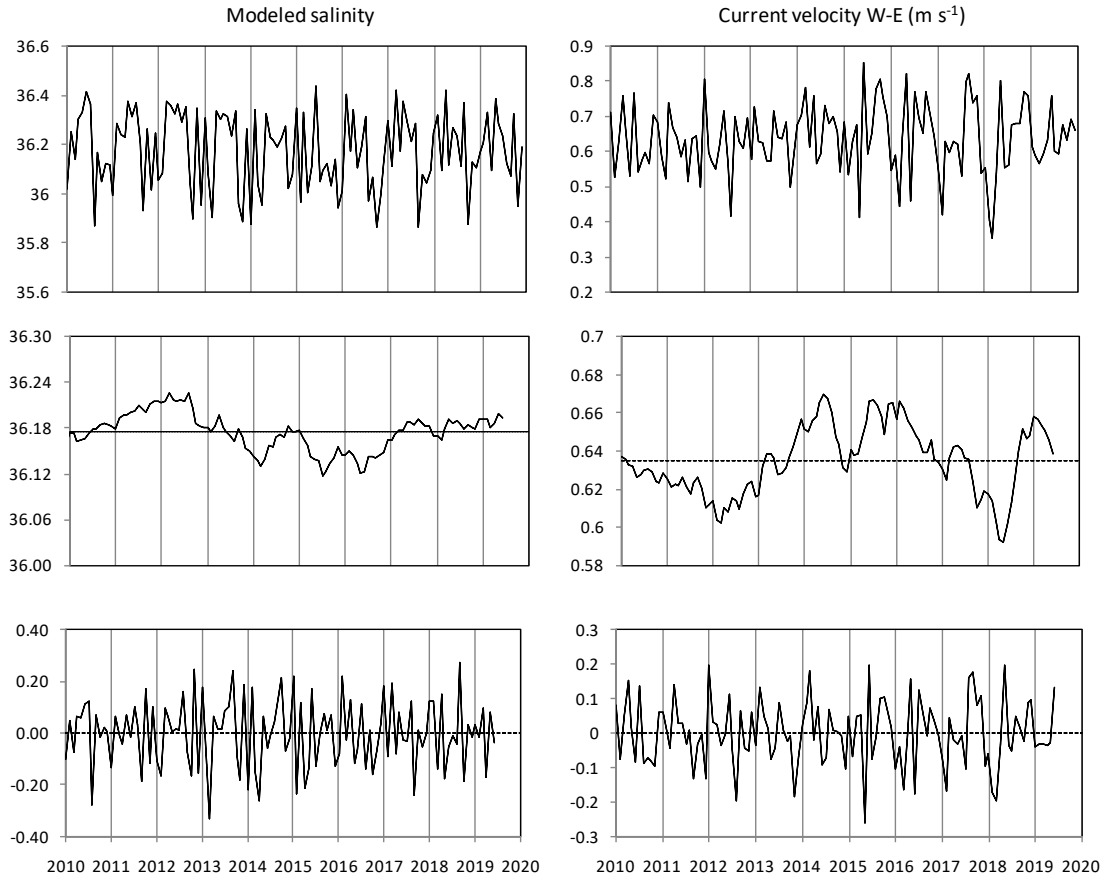
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809 **Fig. 3.** Time series of salinity and West-East current velocity obtained from the model
810 IBI-MFC (calculated from data from CMENS-Copernicus). Upper panels show the raw
811 time series. Middle panels represent the trends (horizontal dashed line is the mean of the
812 trend series). Lower panels indicate the irregular component time series.
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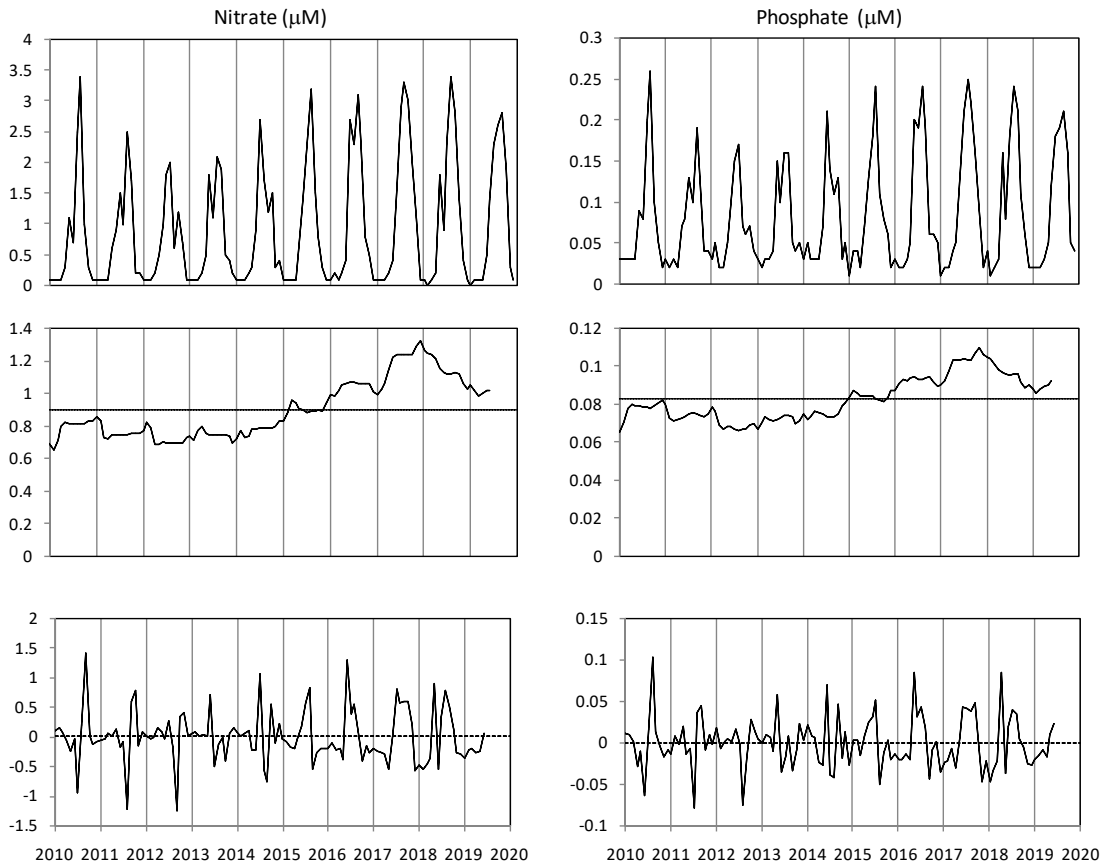


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817 **Fig. 4.** Time series of nitrate and phosphate concentrations obtained from the model
818 IBI-MFC for the position 36.00°N, 5.58°W (CMEMS-Copernicus). Upper panels show
819 the raw data series. Middle panels represent the trend series (horizontal dashed line is
820 the mean of the trend series). Lower panels indicate the irregular component time series.
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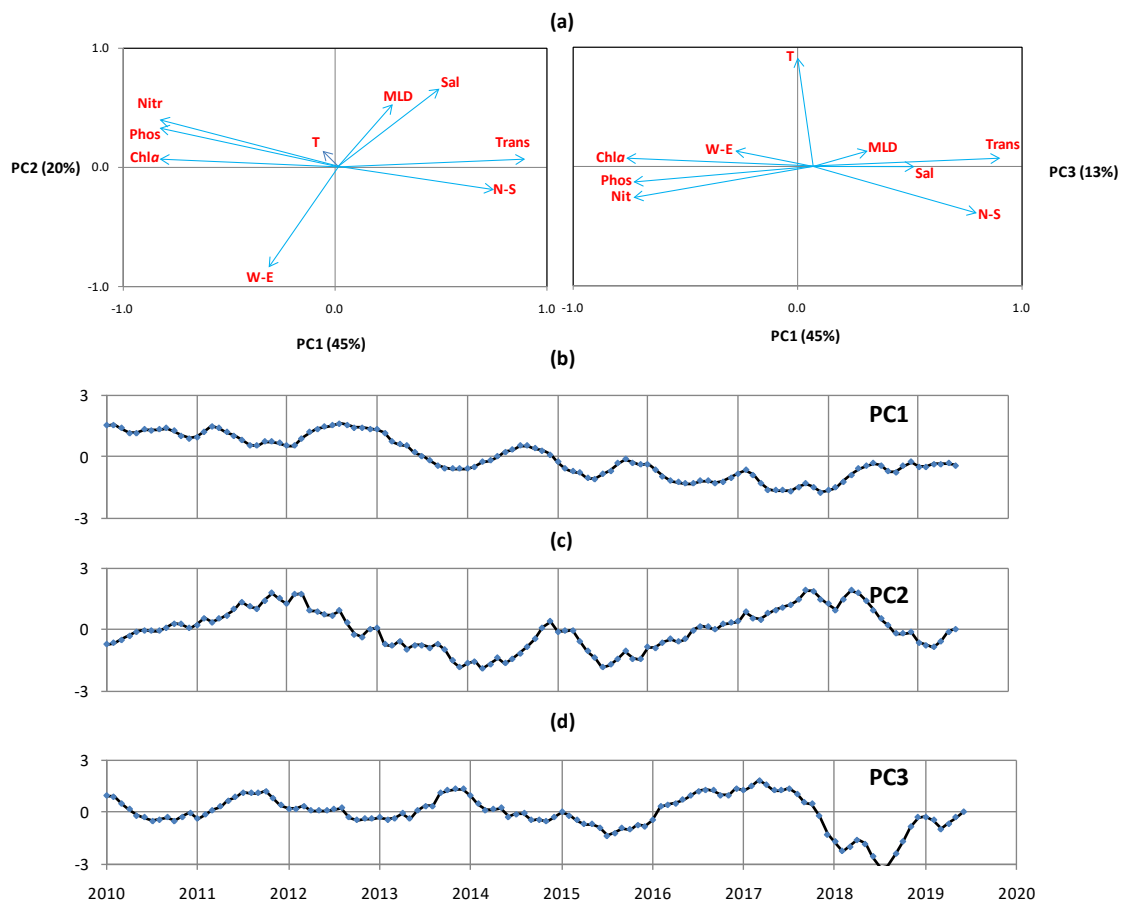
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829 **Fig. 5.** Results of the PCA performed with the modeled variables with the model IBI-
830 MFC (CMEMS-Copernicus) for the position 36.00°N, 5.58°W. **(a)** Biplot of the three
831 first principal components. Variables: Sal, salinity; MLD, mixed layer depth; Nit, nitrate
832 concentration; Phos: phosphate concentration; Chla, chlorophyll *a* concentration; WE,
833 W-E current velocity; T, temperature; Trans, transparency. Time evolution of the three
834 first principal components obtained from the PCA performed with the trend time series
835 of the output variables of the hydro-geochemical model. **(b)**, **(c)** and **(d)** represent the
836 time series of scores for the three first PC.
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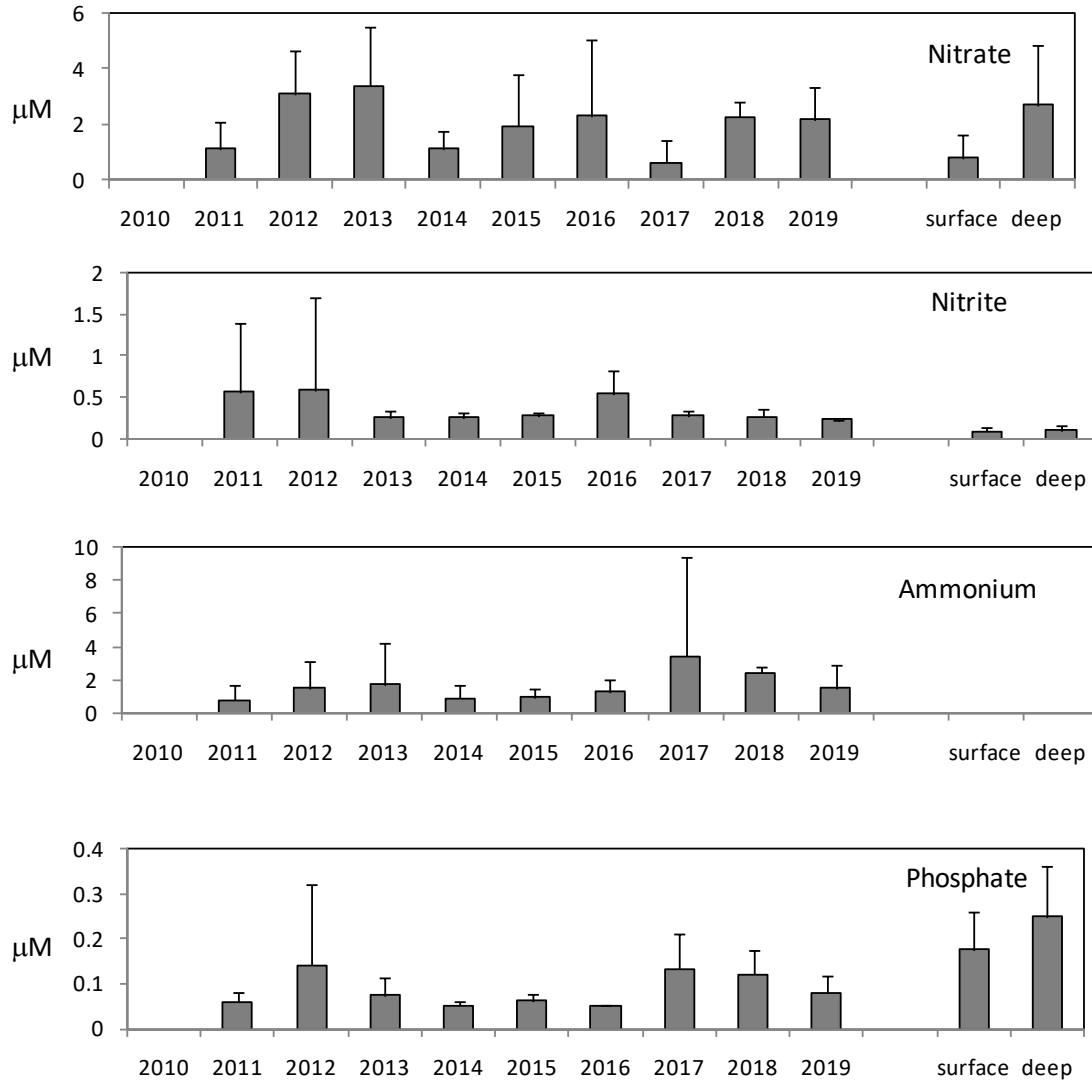
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846 **Fig. 6.** Annual averaged concentrations of nutrients calculated from data obtained at
847 coastal sampling stations (REDIAM, Junta de Andalucía). The data obtained in the open
848 sea station located close to the Bay of Algeciras are also shown for the surface (0-20 m
849 depth) and deep layer (30-100 m depth). The vertical lines indicate the standard
850 deviation.



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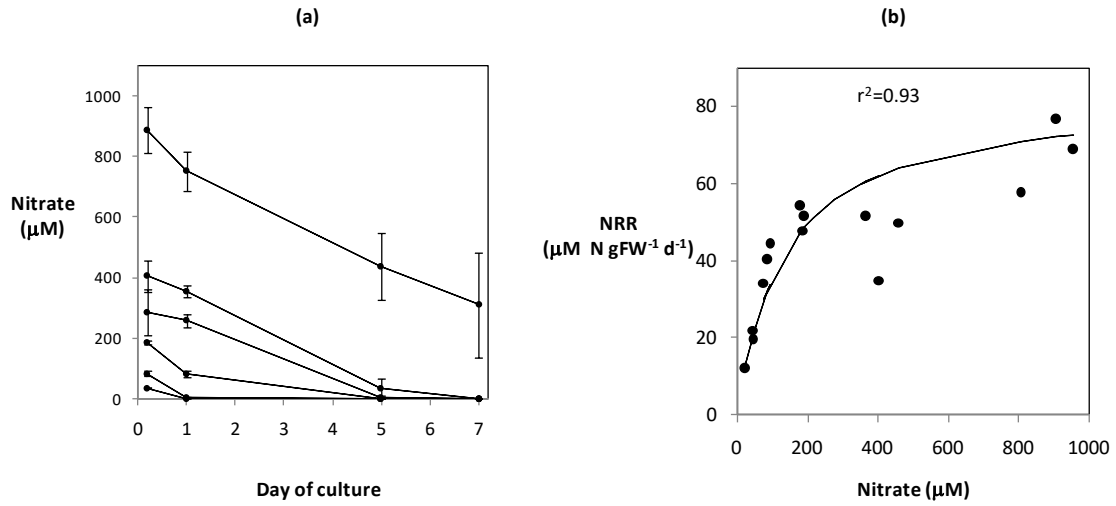
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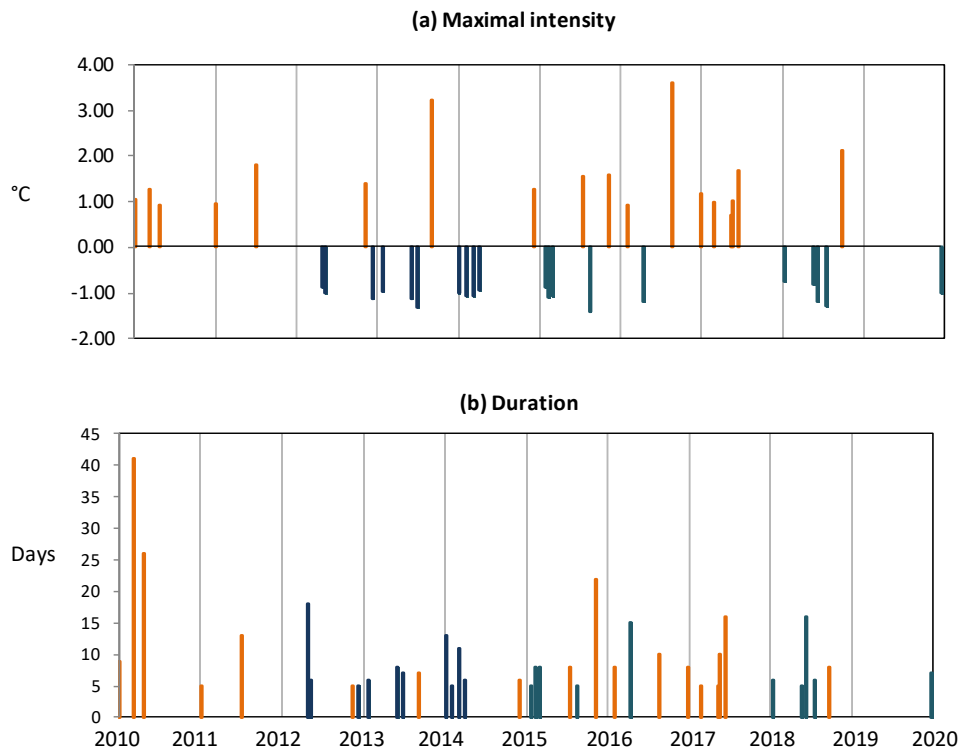
858 **Fig. 7.** Results of the nitrate removal experiments done with *R. okamurae* cultivated
859 under different nutritional conditions. (a) Time course of the nitrate concentration
860 measured in the incubation medium; (b) Nitrate removal rates estimated for the first day
861 of treatment in each incubation vessel (note that each nitrate treatment was assayed by
862 triplicate and that slight differences in the initial concentration were produced at the
863 beginning of the experiments). The continuous line indicates the fitting of the data to the
864 Michaelis-Menten model (r^2 is the coefficient of determination).

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867 **Suppl. Fig. 1.** Time series of heatwaves (red lines) and cold spells (blue lines). The
868 maximal intensity (a) and duration (b) of each extreme temperature event are shown.
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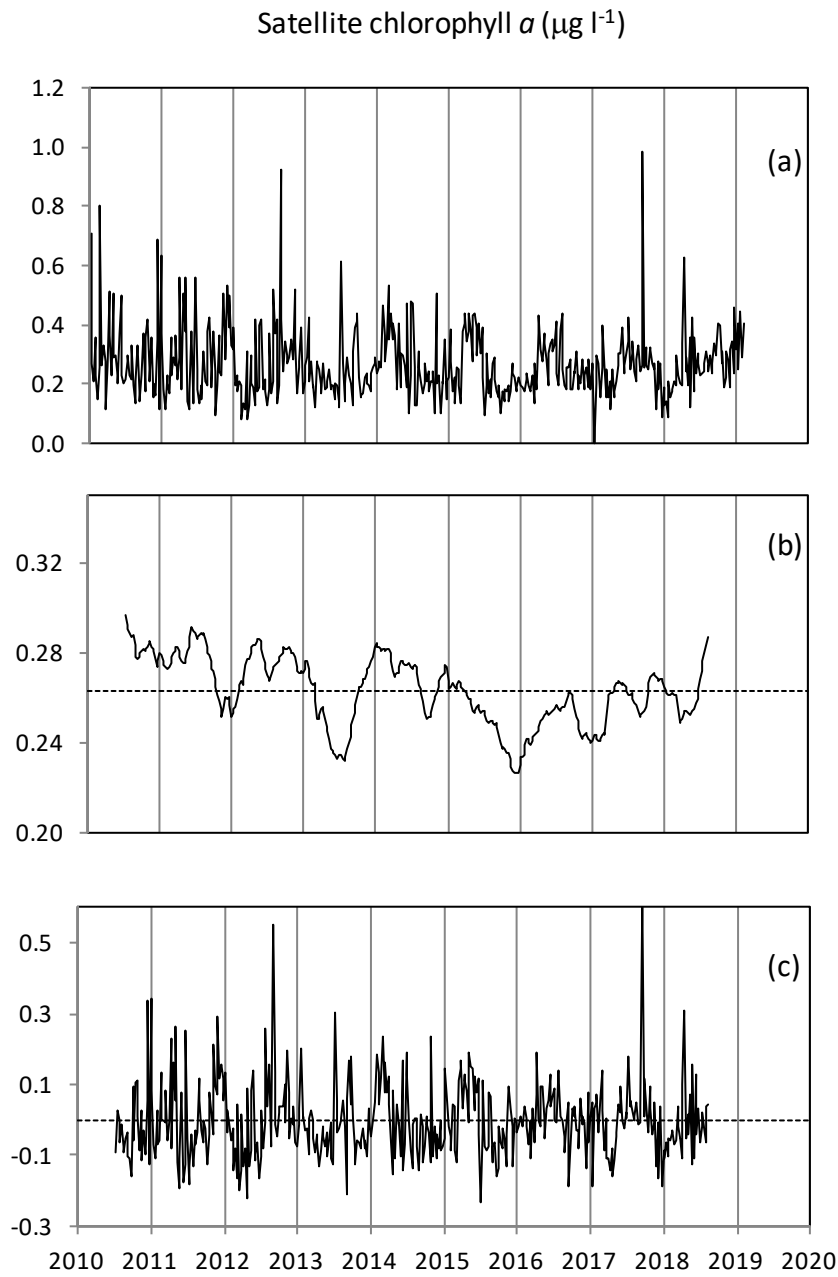
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873 **Suppl. Fig. 2.** Time series of satellite chlorophyll *a*. a) raw data series; b) trend series;
874 c) irregular component series.

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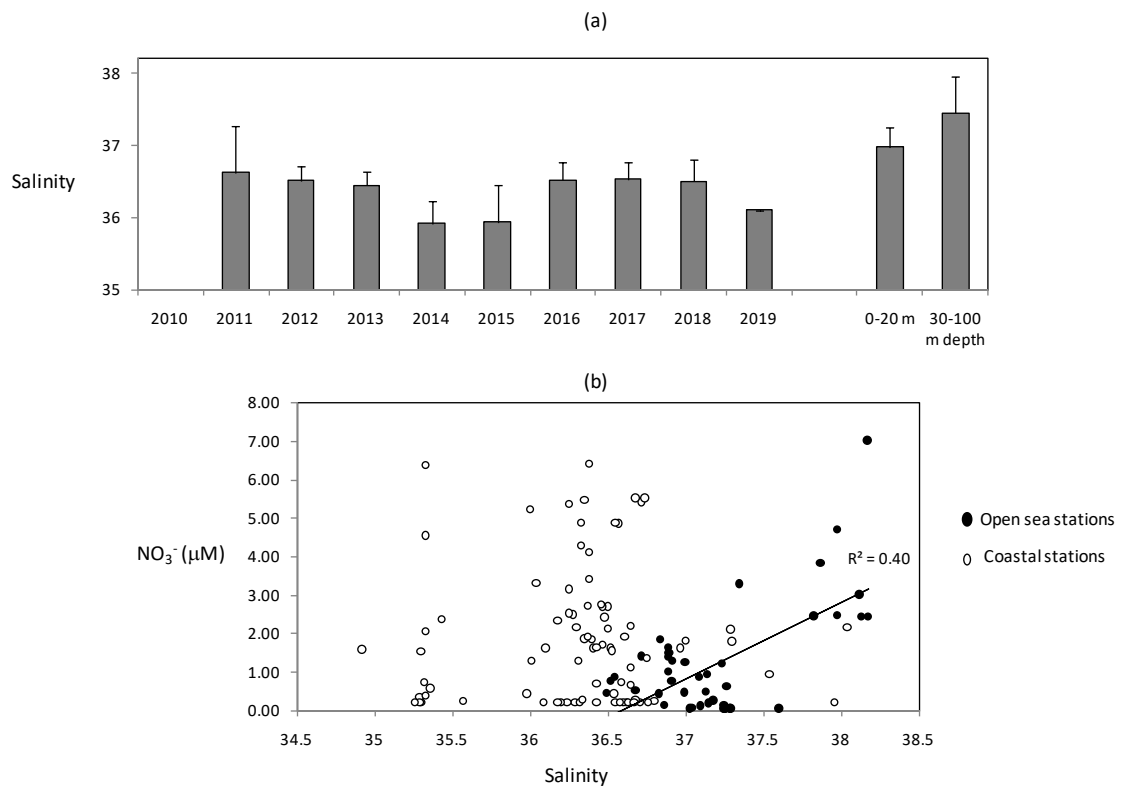
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882 **Suppl. Fig.3.** (a) Annual mean of salinity calculated from data obtained at coastal and
883 open sea sampling stations. For the open sea station, the data for the surface (0-20 m
884 depth) and deep layer (30-100 m depth) are shown. The vertical lines indicate the
885 standard deviation. (b): relationships between salinity and nitrate concentration for the
886 coastal (opened symbol) and open sea (black symbol) stations. Note that the correlation
887 between salinity and nitrate was not statistically significant for the coastal stations.
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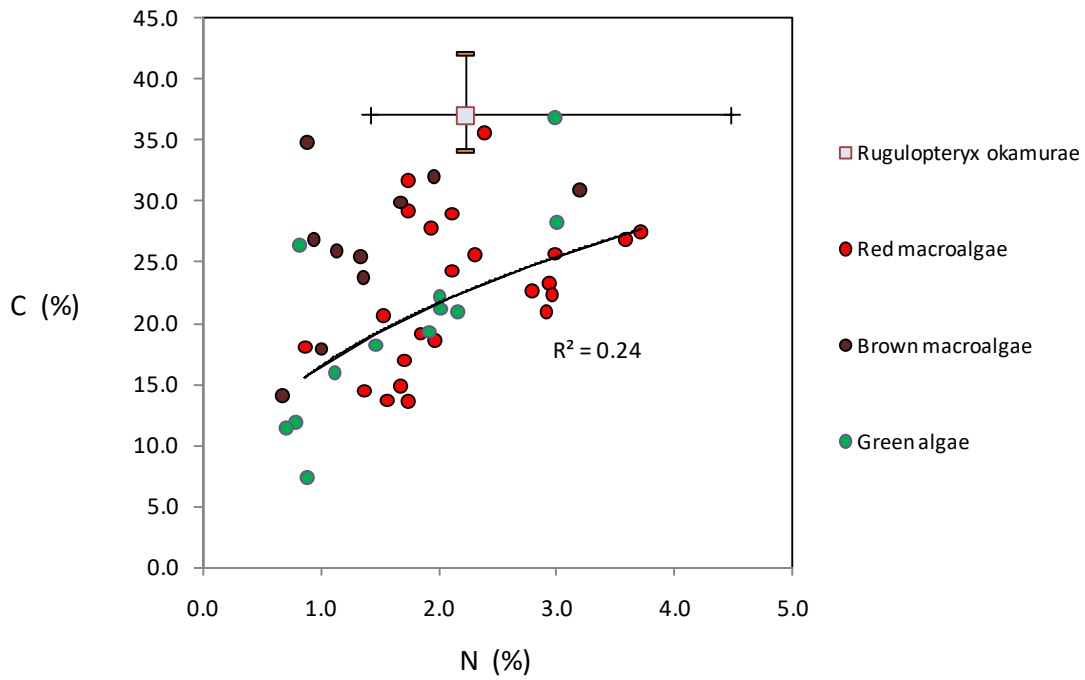
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899 **Suppl. Fig. 4.** Carbon and nitrogen content averaged for thalli of *Rugulopteryx*
900 *okamurae* (open square) collected at different times in 2019. Data obtained for other
901 macroalgae collected in 2004 at the southern Iberian Peninsula are shown. The vertical
902 and horizontal lines indicate the variation ranges of C% and N% obtained for the
903 different thalli of *R. okamurae* that have been analyzed. The data for the other algae
904 were fitted to an exponential curve (the determination coefficient is shown).
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